



Bioaccumulation and biomagnification of potentially toxic elements in the octopus *Octopus hubbsorum* from the Gulf of California



Nefertiti Taydé Roldán-Wong^a, Karen A. Kidd^{b,1}, Ana Judith Marmolejo-Rodríguez^a, Bertha Patricia Ceballos-Vázquez^a, Evgueni Shumilin^a, Marcial Arellano-Martínez^{a,*}

^a Instituto Politécnico Nacional, Centro Interdisciplinario de Ciencias Marinas, Av. Instituto Politécnico Nacional s/n Col. Playa Palo de Santa Rita, C.P. 23096 La Paz, B. C.S., Mexico

^b Canadian Rivers Institute & Biology Department, University of New Brunswick, 100 Tucker Park Road, P.O. Box 5050, Saint John, New Brunswick E2L 4L5, Canada

ARTICLE INFO

Keywords:

Cephalopod
Metals
Ecotoxicology
Human exposure
México

ABSTRACT

The concentrations of 21 potentially toxic elements (PTEs) were determined in the tissues of *Octopus hubbsorum* from three locations along the Gulf of California coast: two near Santa Rosalia (SR), a site with historical metal contamination, and one in La Paz Bay, a reference site. Concentrations of Cd, Co, Cr, Mn, Ni, Pb, and Zn in octopus from the two SR sites were higher than those from the reference site, reflecting the higher sediment concentrations at the mining-impacted locations. The highest bioaccumulation and biomagnification of elements was found in digestive gland and branchial hearts, while the lowest was observed in the mantle, where the mean concentration of PTEs did not exceed international standards for human consumption of octopus. This study found elevated PTEs in octopus from sites with high metal contamination, and presents the first data on these elements in octopus from the Gulf of California.

1. Introduction

Octopus is a high-demand fishery for human consumption throughout the world (FAO, 2016). However, similar to other cephalopods, they have the ability to bioaccumulate high concentrations of potentially toxic elements (PTEs) in their tissues, posing a risk to their consumers (e.g. Napoleão et al., 2005; Rjeibi et al., 2014). For this reason, understanding PTEs in these organisms has become a topic of growing international interest (e.g. Raimundo et al., 2010a; Semedo et al., 2012; Grayson and Sekadende, 2014; Karim et al., 2016).

In marine ecosystems, octopuses play a key role in the transfer of contaminants through food webs (Bustamante et al., 1998a). They feed upon crustaceans and molluscs (FAO, 2016) that accumulate high concentrations of metals and other elements (Jakimska et al., 2011), and are known to accumulate potentially toxic elements (PTEs) such as Cd and As (Seixas et al., 2005a). Octopuses are also part of the diet of many predators including fishes, marine mammals and seabirds, so they act as vectors in the transfer of PTEs to higher trophic levels (Bustamante et al., 1998a, b).

The bioaccumulation of PTEs in octopus occurs differentially in the various organs and tissues, with the digestive gland and branchial

hearts as the main storage and detoxification sites (Nessim and Riad, 2003; Raimundo and Vale, 2008; Raimundo et al., 2010a). Other tissues such as the mantle and arms - the only parts consumed by humans - generally have lower concentrations (e.g. Napoleão et al., 2005; Raimundo and Vale, 2008). However, there are reports that the mantle and arms of octopus exceed the maximum PTE content in international standards for human consumption, posing a risk for human health (Rossi et al., 1993; Grayson and Sekadende, 2014).

In Mexico, the octopus fishery on the Pacific coast is mainly based on the capture of a single species: *Octopus hubbsorum* (López-Uriarte et al., 2005). Within the Gulf of California, the most important location for this fishery is the Santa Rosalia mining port (López-Uriarte et al., 2005), a site known to have potentially toxic levels of metals in coastal marine sediments (Shumilin et al., 2013).

Despite their potential to affect human and ecosystem health, to date there are no studies evaluating PTE concentrations in octopus from Santa Rosalia, nor in any octopus species inhabiting coastal areas of the American continent. The objectives of this research were to determine: (1) the bioaccumulation of PTEs in four tissues of *Octopus hubbsorum*: mantle, branchial hearts, digestive gland, and gills; and (2) the biomagnification of these elements, if any, by analyzing one of its potential

* Corresponding author.

E-mail addresses: ntrw88@gmail.com (N.T. Roldán-Wong), kiddk@unb.ca, karenkidd@mcmaster.ca (K.A. Kidd), amarmole@ipn.mx (A.J. Marmolejo-Rodríguez), bceballo@ipn.mx (B.P. Ceballos-Vázquez), eshumili@ipn.mx (E. Shumilin), marellam@ipn.mx (M. Arellano-Martínez).

¹ Present address: Department of Biology & School of Geography and Earth Sciences, McMaster University, 1280 Main Street West, Hamilton, Ontario, Canada, L8S 4K1.

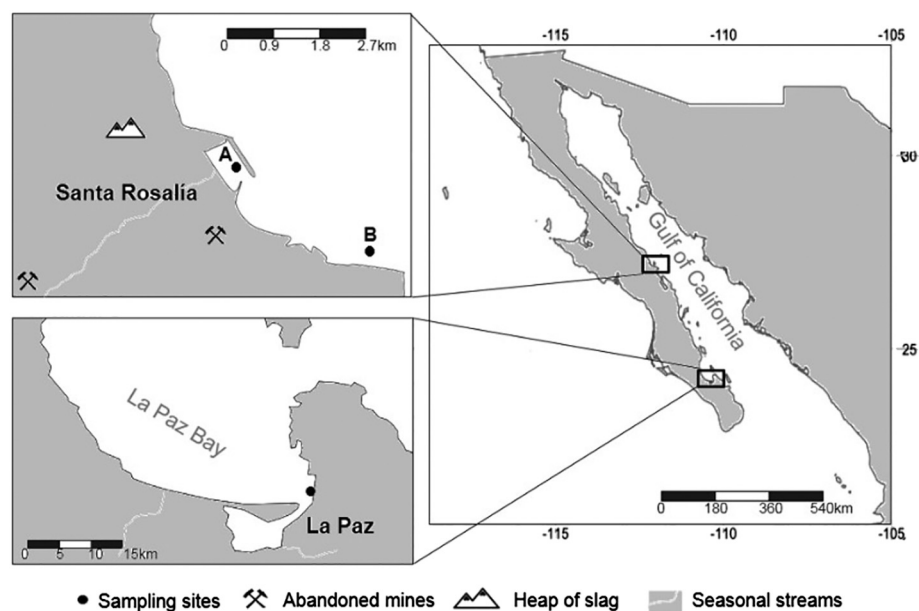


Fig. 1. Sampling sites of *Octopus hubbsorum* on the Gulf of California: two sites near Santa Rosalía (A and B) with historical metal contamination in marine sediments, and one in La Paz Bay, as a reference site.

prey, the clam *Megapitaria squalida*, collected from the Santa Rosalía mining port. Results from the mining-impacted sites were compared to the same species from La Paz Bay, BCS, Mexico, a reference site.

2. Materials and methods

2.1. Study area

Santa Rosalía is a fishing port located in the central part of the Gulf of California's western coast, Mexico (Fig. 1). The local climate is arid, with a mean temperature of 23.4 °C, and scarce rainfall, with annual mean precipitation of 117.2 mm (Volke-Sepúlveda et al., 2003). It is a location with a long history of mining activity related to copper extraction from “El Boleo” field (1885–1985), so that the local coastal sediments are highly polluted by metals, mainly Mn, Cu, Zn, Pb, Co, U and Cd (Shumilin et al., 2013). The concentrations of these elements are so high that they are potentially toxic for marine biota (Shumilin et al., 2013). Specifically, Cu concentrations exceeded the effect range medium (ERM) by up to 17-fold, which translates into a potential negative biological effect for 50% of marine organisms (Shumilin et al., 2011). The local mussels *Modiolus capax* (Gutiérrez-Galindo et al., 1999; Muñoz-Barbosa and Huerta-Díaz, 2013) and *Mytilus edulis* (Cadena-Cárdenas et al., 2009) have Cu concentrations up to 11 times higher than those from other areas of the Baja California peninsula; both bivalve species are part of the diet of *O. hubbsorum* (López-Uriarte et al., 2010).

2.2. Sampling

Cd, Co, Cu, Mn, Pb, and Zn levels in Santa Rosalía marine sediments show a marked decreasing gradient toward the port's most remote areas (Shumilin et al., 2011, 2013). Accordingly, sampling was conducted in two locations: at (A) the area closest to the port dock (Hot spot zone), where the highest sediment concentrations of these elements have been recorded, and at (B) an area located 2 km to the south, where lower sediment concentrations have been found (Fig. 1). In addition, reference samples were collected from El Caimancito, La Paz Bay. This site is 400 km south of Santa Rosalía and sediment metal concentrations here are related mainly to the regional geology and deemed typical of pristine environments with no anthropogenic impact; the effects of the closest mining area (the mining company Roca Fosfórica Mexicana “Rofomex”) are localized (Rodríguez-Castañeda et al., 2006) (Fig. 1).

Twenty three *O. hubbsorum* specimens were collected from the following locations: Santa Rosalía A (n = 8), Santa Rosalía B (n = 7) and La Paz Bay (n = 8). Specimens were caught by artisanal fishermen in October 2015 and June 2016. To avoid any potential bias from individuals of extreme size or weight, all octopuses were within a similar range of size and weight. Specimens were transported in sea water with ice to reduce their metabolic rate and minimize the mobilization of PTEs between organs and tissues (Martin and Flegal, 1975). Immediately after being killed, the sex, total weight (TW), total length (TL), and dorsal mantle length (DML) were determined. The mantle (mantle and arms are the main tissues consumed by humans but mantle is more frequently used in similar studies), digestive gland (main storage tissue), branchial hearts (main detoxification tissue) and gills (tissue with major interaction with water and exposed to the water-borne contamination) of each specimen was removed. Each tissue was carefully washed with deionized water, weighed (± 0.1 g), frozen (-20 °C), freeze-dried (-30 °C and 100 m Torr), powdered and homogenized in agate mortars for analysis. To minimize the risk of contamination, organisms and samples were handled only with acrylic and plastic materials previously washed with deionized water and Milli-Q(R) metal-free water with 5% nitric acid. Analyses were performed in the Canadian Rivers Institute at the University of New Brunswick, NB, Saint John, Canada.

2.3. Determination of potentially toxic elements

Using between 0.001 and 0.02 g (dry weight) of each sample, total Hg (THg) was determined by thermal decomposition, amalgamation and thermal absorption, with a Milestone DMA-80 direct mercury analyzer. Quality assurance was assessed every 10 samples using a blank (< 0.00250 mg kg⁻¹ limit of detection), a duplicate (relative percent difference $< 20\%$), a mussel tissue certified reference material SRM 2976 and an internal reference material (recovery percentages between 80 and 120%). The limit of detection (LOD) was determined using $3 \times$ the SD of the blanks run over the project plus the average of the blanks (see Table 1).

For the determination of Ag, Al, As, Cd, Co, Cr, Cu, Fe, La, Mg, Mn, Ni, Pb, Rb, Se, Sr, Tl, U, V and Zn, approximately 0.5 g (dry weight) of each sample was digested with 10 mL of concentrated nitric acid (HNO₃) in a CEM MARS5 microwave digester for approximately 1 h. Then, each sample was diluted in 40 mL Milli-Q(R) metal-free water, and was analyzed by inductively coupled plasma optical emission

Table 1

Detection limits, concentrations of potentially toxic elements in the certified reference material SRM 2976 (mean \pm SD) and percentage recovery according the concentrations obtained.

Element	Detection limit (mg kg ⁻¹ , dry weight)	SRM 2976 (mg kg ⁻¹ , dry weight)	Recovery (%)
Ag	0.22	0.011 \pm 0.005	98
Al	0.31	134 \pm 34	111
As	1.60	13.3 \pm 1.8	103
Cd	0.13	0.82 \pm 0.16	107
Co	0.15	0.61 \pm 0.02	89
Cr	0.12	0.50 \pm 0.16	108
Cu	0.19	4.02 \pm 0.33	96
Fe	0.68	171.0 \pm 4.9	97
Hg	0.01	61.0 \pm 3.6	79
La	1.00	n. c.	–
Mg	4.25	0.53 \pm 0.05	92
Mn	0.10	33.0 \pm 2.0	108
Ni	0.13	0.93 \pm 0.12	98
Pb	0.77	1.19 \pm 0.18	104
Rb	0.86	4.14 \pm 0.09	93
Se	1.08	1.80 \pm 0.15	105
Sr	0.004	93.0 \pm 2.0	84
Tl	0.55	n. c.	–
U	6.37	n. c.	–
V	0.15	n. c.	–
Zn	1.02	137 \pm 13	97

n. c. - no certified.

spectrometry (ICP-OES), using 1.5 mL of 29% lithium nitrate as an ionization buffer and 100 μ L of yttrium (Y, 1000 mg/L) as the internal standard. The accuracy and precision were determined from reading the point of calibration (1.0 and 20.0 ppm) before and after each batch of samples (recovery percentage between 80 and 120%). A batch of samples consisted of method blank (\leq detection limit), 12 samples, a duplicate (relative percent difference $<$ 20%) and mussel tissue as a certified reference material SRM 2976 (recovery percentage between 79 and 111%). The detection limit (DL) was calculated as 5 times the instrument detection limit (IDL); the IDL was determined by running 20 repeats of a blank and then adding 3 times the SD to the average (see Table 1).

The concentrations of each element including Hg are reported in mg kg⁻¹ dry weight.

2.4. Analyses of data

Considering that the digestive gland is the main site of PTE bioaccumulation in cephalopods (Miramand and Guary, 1980), the extent to which accumulation in this tissue was greater than mantle muscle was determined. To this end, we used the classification proposed by Miramand and Bentley (1992), where the ratio of the concentration of each element in the digestive gland relative to the mantle (DG:M) indicates whether the element is highly concentrated (ratio $>$ 50), moderately concentrated (50 $>$ ratio $>$ 10) or slightly concentrated (ratio $<$ 10). In addition, since branchial hearts are also major PTE bioaccumulation sites (Mangold et al., 1989), the same classification was applied to determine the extent to which PTE concentration in this organ was greater than that of the mantle (BH:M).

To evaluate biomagnification in *O. hubbsorum*, we used PTE concentrations in the clam *Megapitaria squalida*, a common prey in the diet of this octopus (personal observation) as well as in the diet of other local octopuses such as *Octopus bimaculatus* (Armendáriz-Villegas et al., 2014). Data were obtained from sampling conducted in 2013 at three similar locations: Santa Rosalia port (an area with a high metal concentrations) (n = 4), 7 km north of Santa Rosalia (an area with lower metal concentrations) (n = 4), and La Paz Bay (n = 2). The analyses were done on whole body clams, using the same analytical procedures described above. Biomagnification was assessed by calculating the

Table 2

Size and weight (mean \pm SD, range) of *Octopus hubbsorum* from three locations in the Gulf of California.

	Santa Rosalía A	Santa Rosalía B	La Paz Bay	ANOVA
Dorsal mantle length (cm)	8.6 \pm 1.4 (7–11)	8.6 \pm 1.3 (7–11)	9.0 \pm 1.5 (6.5–11)	F _{2,23} = 0.25, p = 0.78
Total length (cm)	35.5 \pm 4.4 (28–39)	34.9 \pm 2.8 (32–40)	37.1 \pm 6.3 (28–44)	F _{2,23} = 0.46, p = 0.64
Total weight (g)	375 \pm 167 (228–673)	438 \pm 120 (270–577)	478 \pm 157 (286–710)	F _{2,23} = 0.95, p = 0.40

biomagnification factor (BMF) for each site, as follows: PTE concentration in *O. hubbsorum* (C₂), relative to the mean concentration in the prey *M. squalida* (C₁), expressed as BMF = C₂/C₁ (Newman, 2014).

Any potential differences in PTE concentrations between sexes, organs and localities were examined using one-way analyses of variance (ANOVA) and Kruskal-Wallis nonparametric tests, according to compliance with the assumptions of normality and homogeneity of variances, checked through the Kolmogorov-Smirnov and Bartlett tests, respectively. To determine the source of variation, a posteriori Tukey and Bonferroni tests were used after ANOVA and Kruskal-Wallis tests, respectively. All statistical analyses were performed in the software STATISTICA for Windows (version 12.0, Stat Soft 2013), with a 95% confidence level (α = 0.05).

3. Results

The size and weight of the octopuses examined ranged from 7 to 11 cm mantle length, 28 to 44 cm total length, and 228 to 710 g total weight. La Paz Bay specimens displayed the highest values of all measurements; however, no significant differences were found in size and weight between the three sampling localities (ANOVA, p $>$ 0.05) (Table 2).

3.1. Concentration of potentially toxic elements in tissues

The metal concentrations that were determined in octopus from these sites included nine elements in the mantle (Al, As, Cu, Fe, Hg, Mg, Mn, Sr and Zn); 17 in the digestive gland (Ag, Al, As, Cd, Co, Cr, Cu, Fe, Hg, Mg, Mn, Ni, Pb, Sr, V and Zn); 16 in the branchial hearts (Al, As, Cd, Co, Cr, Cu, Fe, Hg, Mg, Mn, Ni, Se, Sr, U, V and Zn); and 14 in the gills (Al, As, Cd, Co, Cr, Cu, Fe, Hg, Mg, Mn, Ni, Se, Sr, and Zn). It was not possible to determine the concentration of all 21 elements in each tissue as some were below the detection limit. Ag and Pb were detected only in the digestive gland, and U was found only in the branchial hearts. La, Rb and Tl were below the detection limits in all tissues ($<$ 1.00, $<$ 0.86 and $<$ 0.55 mg kg⁻¹, respectively). The mean concentrations and ranges for each element by site and tissue are shown in Table 3.

3.2. Effect of sex on the concentration of potentially toxic elements

Of the 23 octopus specimens examined, 39% were females and 61% were males, with a female: male sex ratio of 1:1.5; the ratio by site was 1:3 (Santa Rosalia A), 1:1.3 (Santa Rosalia B) and 1:1 (La Paz Bay). PTE concentrations in the tissues showed no significant differences between males and females in each site (p $>$ 0.05), suggesting that sex did not affect PTE accumulation in organisms within the size/weight ranges examined. Thus, the data from both sexes were pooled for subsequent analyses.

3.3. Differences in PTE concentrations between tissues

A comparison of PTE concentrations between octopus tissues across

Table 3
Concentration of potentially toxic elements (mg kg⁻¹, dry weight) in mantle, digestive gland, branchial hearts and gills of *Octopus hubbsorum* and other octopus species.

Tissue	Species	Locality	Ag	Al	As	Cd	Co	Cr	Cu	Fe	Hg	La	Reference	
Muscle (mantle) *Mantle and arms **Arms	<i>Octopus hubbsorum</i>	Santa Rosalia	< 0.22	3.7 (1.1–7.5)	37 (14–55)	< 0.13	< 0.15	< 0.12	25 (6–37)	10 (1–20)	0.07 (0.02–0.25)	< 1.00	Present study	
		La Paz Bay	< 0.22	5.6 (0.5–11)	65 (25–93)	< 0.13	< 0.15	< 0.12	20 (14–29)	9 (5–15)	0.13 (0.07–0.24)	< 1.00	Present study	
Digestive gland	<i>Octopus vulgaris</i>	Portugal				0.04–3.3			3.9–72	6.5–81			Raimundo et al., 2004	
		Portugal*			13–87				5.5–72	24 ± 15			Raimundo et al., 2005	
		Portugal*				23			30 ± 19	32 (11–84)			Napoleão et al., 2005	
		Affric				0.08 ± 0.04			36 (17–106)	30 ± 5			Soldevilla, 1987	
		Monaco				0.25–4.4			26 ± 2	17–43			Miramand and Guary, 1980	
		Portugal							8.3–62				Raimundo and Vale, 2008	
		Portugal**			40–133							0.27–0.48		Seixas et al., 2005b
		Egypt				0.4–19				7.9–81	15–49	0.15–0.43		Seixas et al., 2005a
		Tunisia				1.3–1.8			1.7–1.9	5.8–5.9	10.5–15.6			Nessim and Riad, 2003
		Morocco				0.15–0.5			1.2–2.7		40.6–229	0.1–0.4		Rjeibi et al., 2014
		Indian Ocean				0.24–0.58				133–171				Karim et al., 2016
		Indian Ocean				0.37				15				Bustamante et al., 1998b
France				0.21				3				Bustamante et al., 1998b		
Italy				0.24 ± 0.01				17 ± 1	25 ± 10			Miramand and Bentley, 1992		
Branchial hearts	<i>Octopus hubbsorum</i>	Santa Rosalia	0.05	8	43	76	50	0.24	3296	314	0.5–9.0	< 1.00	Rossi et al., 1993	
		La Paz Bay				(0.3–1.3)			(21–252)	(162–644)	0.29 (0.07–1.06)		Present study	
		Portugal				2.7			0.49	2104	454	0.22		Present study
		Portugal				(1.6–3.1)			(0.2–0.7)	(537–4800)	(250–878)	(0.15–0.27)		
		Egypt								137–3142				Raimundo et al., 2004
		Tunisia								137–1465	292–1202			Raimundo et al., 2005
		Morocco								1768 ± 1010	790 ± 343			Napoleão et al., 2005
		Indian Ocean				16–81				82 (8.6–286)	456			Soldevilla, 1987
		Indian Ocean								2500 ± 700	700 ± 130			Miramand and Guary, 1980
		France								4200				Raimundo et al., 2010a
		Santa Rosalia								28–762	142–384	0.58–3.43		Raimundo and Vale, 2008
		Gills	<i>Octopus vulgaris</i>	Portugal							390–715	138–179		
Monaco									800–1000		0.5–3		Nessim and Riad, 2003	
Portugal									152–187	104–344			Rjeibi et al., 2014	
Portugal									1092				Karim et al., 2016	
Egypt									42				Bustamante et al., 1998b	
Indian Ocean									436–483	255–288			Bustamante et al., 1998b	
Indian Ocean									4	1028			Miramand and Bentley, 1992	
France									(1–7)	(156–5340)			Present study	
Santa Rosalia									2	618				Present study
La Paz Bay									(1–5)	(300–1580)				Present study
Portugal										188 ± 68	577 ± 323			Napoleão et al., 2005
Monaco										500 ± 40	650 ± 150			Miramand and Guary, 1980
Portugal								92–274	95–1270			Raimundo and Vale, 2008		
Portugal								57–121	40.8–69.3	0.27–0.52		Seixas et al., 2005b		
Egypt								465				Nessim and Riad, 2003		
Indian Ocean								306				Bustamante et al., 1998b		
Indian Ocean								335 ± 14	21 ± 2			Bustamante et al., 1998b		
France								136	91			Miramand and Bentley, 1992		
Santa Rosalia								(84–187)	(16–529)	0.05 (0.02–0.11)		Present study		
La Paz Bay								130	44	0.08 (0.04–0.14)		Present study		
Portugal								(73–174)	(30–65)			Napoleão et al., 2005		

(continued on next page)

Table 3 (continued)

Tissue	Species	Locality	Ag	Al	As	Cd	Co	Cr	Cu	Fe	Hg	Ia	Reference
Muscle (mantle) **Mantle and arms **Arms Muscle (mantle)	<i>Octopus hubbsorum</i>	Affric	17						177(92–253)	49 (19–87)			Soldevilla, 1987
		Monaco	0.05 ± 0.01						300 ± 40	40 ± 20			Miramand and Guary, 1980
		Portugal	0.19–0.90										Raimundo et al., 2010a
		Portugal	0.16–22						66–128	10–252	0.28–0.42		Raimundo and Vale, 2008
		Egypt	1.2–1.9										Seixas et al., 2005b
		Morocco	0.37–0.58						23–38	10.8–15.8			Nessim and Riad, 2003
		Indian Ocean	22.6						371–644	208–407			Karim et al., 2016
		Indian Ocean	49.1						530				Bustamante et al., 1998b
		France	0.56						168				Bustamante et al., 1998b
			0.11						268 ± 8	9 ± 3			
Digestive gland	<i>Octopus vulgaris</i>	Portugal	2741 (923–3770)	1.3 (0.3–2.6)	< 0.13	< 0.77	< 0.86	< 1.08	13 (2–19)	< 0.55	< 0.15	58 (19–74)	Present study
		La Paz Bay	3164 (2420–4910)	0.9 (0.7–1.1)	< 0.13	< 0.77	< 0.86	< 1.08	14 (9–28)	< 0.55	< 0.15	64 (58–73)	Present study
		Portugal				0.06–2.3							
		Portugal*											Raimundo et al., 2004
		Portugal*											Raimundo et al., 2005
		Affric											Napoleão et al., 2005
		Monaco											Soldevilla, 1987
		Portugal											Miramand and Guary, 1980
		Portugal**											Raimundo and Vale, 2008
		Egypt											Seixas et al., 2005a
Branchial hearts	<i>Octopus vulgaris</i>	Tunisia											Nessim and Riad, 2003
		Morocco											Rjebi et al., 2014
		Tanzania*											Karim et al., 2016
		Indian Ocean											Mshana and Sekakende, 2014
		Indian Ocean											Bustamante et al., 1998b
		France											Bustamante et al., 1998b
		Santa Rosalia											Miramand and Bentley, 1992
		La Paz Bay											Present study
		Portugal											Present study
		Portugal											Present study
Digestive gland	<i>Octopus vulgaris</i>	Santa Rosalia	1893 (1300–2630)	1.0 ± 0.3 (3–37)	0.4 ± 0.1 (0.5–2.8)	0.11 ± 0.05 (1–14)	< 5.0 (4–19)	8 (5–10)	21 (16–31)	< 6.37 (0.5–4.3)	< 0.5	877 (226–1500)	Present study
		La Paz Bay	2014 (1300–2620)	6 (3–7)	2.4 (1–4.5)	2.7 (0.7–5.9)	< 5.0 (5–10)	32 (16–48)	< 5.0 (0.5–2.7)	< 6.37 (0.5–2.7)	2.0 (468–1190)	802 (198–14,718)	Present study
		Portugal											Raimundo et al., 2004
		Portugal											Raimundo et al., 2005
		Portugal											Napoleão et al., 2005
		Affric											Seixas and Pierce, 2005
		Monaco											Soldevilla, 1987
		Portugal											Miramand and Guary, 1980
		Portugal											Raimundo et al., 2010a
		Egypt											Raimundo and Vale, 2008
Branchial hearts	<i>Octopus vulgaris</i>	Tunisia											Nessim and Riad, 2003
		Morocco											Rjebi et al., 2014
		Tanzania											Karim et al., 2016
		Indian Ocean											Mshana and Sekakende, 2014
		Indian Ocean											Bustamante et al., 1998b
		France											Bustamante et al., 1998b
		Santa Rosalia											Miramand and Bentley, 1992
		La Paz Bay											Present study
		Portugal											Present study
		Portugal											Present study

(continued on next page)

Table 3 (continued)

Tissue	Species	Locality	Mg	Mn	Ni	Pb	Rb	Se	Sr	Tl	U	V	Zn	Reference
Gills	<i>Octopus vulgaris</i>	Portugal		7.8 ± 4.8	251 ± 122	8.1 ± 5.0	7–12	5.2–13.3				36 ± 17 17–78	81 ± 23	Napoleão et al., 2005
		Portugal				0.39–3.9							68–386	Seixas and Pierce, 2005
	<i>Granelledone</i> sp.	Portugal			13 ± 3							25 ± 2	65 ± 15	Raimundo and Vale, 2008
		Monaco			0.9–2.8	34.2–74.4	8.27–9.41						13–58	Miramand and Guary, 1980
	<i>Benthooctopus thielei</i>	Egypt											126	Nessim and Riad, 2003
		Indian Ocean											172	Bustamante et al., 1998b
	<i>Elledone cirrhosa</i>	Indian Ocean			4.5 ± 0.2	6.1 ± 0.2	0.39 ± 0.04						6.0 ± 1.0	Bustamante et al., 1998b
		France			7	0.8	< 0.77	< 5.0	4	14	< 0.5	< 6.37	< 0.15	87
	<i>Octopus hubbsorum</i>	Santa Rosalia		2489	(1620–2850)	(0.3–2.8)			(3–6)	(9–18)				Present study
		La Paz Bay		2530	6	0.3	< 0.77	< 5.0	4	16	< 0.5	< 6.37	< 0.15	80
<i>Octopus vulgaris</i>	Portugal		(2060–3050)	(4–6)	(0.2–0.5)			(3–6)	(11–23)				(74–96)	
		Affric		6.3 ± 2.3	1.5 ± 0.5									72 ± 17
	Monaco			7.6									68 (44–94)	Soldevilla, 1987
		Portugal		10 ± 0.5								0.5 ± 0.2	120 ± 20	Miramand and Guary, 1980
	Portugal					0.2–1.1							726–912	Raimundo et al., 2010a
		Egypt			0.1–1.1	5.5–8.4	2.1–4.4						76–122	Raimundo and Vale, 2008
	Morocco					1.57–7.6							10–20	Nessim and Riad, 2003
		Indian Ocean											56–461	Karim et al., 2016
	<i>Benthooctopus thielei</i>	Indian Ocean											98	Bustamante et al., 1998b
		France			4.3 ± 0.1	0.7 ± 0.1	0.62 ± 0.01						< 0.5	147
<i>Elledone cirrhosa</i>												133 ± 1	Miramand and Bentley, 1992	

all sites revealed general patterns in their accumulation (Table 3). The digestive gland had the highest concentrations of Cd (21–252 mg kg⁻¹), Cu (537–5880 mg kg⁻¹), Hg (0.07–1.06 mg kg⁻¹), Pb (0.7–14 mg kg⁻¹), and Zn (226–1500 mg kg⁻¹) (p < 0.05), with mean values being three (Hg) to 122 (Cu) times higher than those in all other tissues. Branchial hearts showed the highest levels of Al (14–62 mg kg⁻¹), Co (12–597 mg kg⁻¹), Cr (1–7 mg kg⁻¹), Mn (5–115 mg kg⁻¹), Ni (73–364 mg kg⁻¹), Sr (19–93 mg kg⁻¹) and V (13–55 mg kg⁻¹) (p < 0.05), with mean concentrations being two (Sr) to 358 (Co) times higher than in the other tissues. The mantle was the tissue with the lowest concentrations of all the elements (p < 0.05), with mean values from one time (As, Hg, Mg) up to 122 times (Cu) lower than those in the other tissues. Element concentrations in gills were similar to (Al, As, Co, Cr, Cu, Fe, Hg, Mg, Mn, Ni, Sr, Zn) or higher (Cd, Se) than those in the mantle (one to seven-fold), and always lower than those in the digestive gland and branchial hearts.

In the digestive gland of *O. hubbsorum* from all localities, Cd and Cu were highly concentrated (DG:M ratio = 524 and 129, respectively), Zn was moderately concentrated (DG:M ratio = 16), and Pb and Hg were slightly concentrated (DG:M ratio = 6 and 3, respectively), as per the classification of Miramand and Bentley (1992). With regards to the branchial hearts, Co, Ni and V were highly concentrated (BH:M ratio = 535, 1712 and 257, respectively), Al, Cr and Mn were moderately concentrated (BH:M ratio = 11, 32 and 41, respectively), and Sr was slightly concentrated (BH:M ratio = 6).

3.4. Differences in PTE concentrations between localities

There were no significant differences in PTEs within octopus tissues from the two Santa Rosalia localities (p > 0.05). However, in one or both localities (A and/or B) some organs had concentrations (Cd, Co, Cr, Mn, Ni, Pb and Zn) that were significantly higher than those in the same tissues from La Paz Bay. These differences were found in the digestive gland (Co, Mn, and Pb), branchial hearts (Cd, Co, Cr, Mn and Ni) and gills (Cd, Cr, Mn and Zn) (p < 0.05), but never in the mantle (p > 0.05) (Figs. 2 and 3). In contrast, specimens from La Paz Bay showed significantly higher concentrations of Ag, As, Cr, Hg, Ni and Sr when compared to one or both Santa Rosalia sites. More specifically, differences were found in at least one of the organs: mantle (As and Hg), digestive gland (Ag, Cr, Ni and Sr), branchial hearts (As and Hg) and gills (As) (p < 0.05) (Figs. 2 and 3).

With few exceptions, the ratios of most PTEs between digestive gland-mantle and branchial hearts-mantle were similar among sites, according to the classification of Miramand and Bentley (1992). However, in Santa Rosalia Co was more concentrated in the digestive gland and Mn in the digestive gland and branchial hearts; in La Paz Bay, Ag and Al were more concentrated in the digestive gland.

3.5. Estimation of biomagnification

According to the BMFs, all *O. hubbsorum* organs showed some degree of biomagnification, mainly in Santa Rosalia specimens (Table 4). Some elements (Cd, Co, Cu, Fe, Mn, Ni, Pb and Zn) in the digestive gland and branchial hearts of *O. hubbsorum* were 1 to 300 times higher than those of whole *M. squialida*. The mantle also had BMFs > 1 but with a smaller number of elements (Cu, Mn and Zn); concentrations of these PTEs were 1 to 3 times higher than those recorded in whole *M. squialida*. In addition, biomagnification of some elements was also evident when concentrations in *O. hubbsorum* were compared with literature reports of PTEs for other molluscs from the same localities (Table 5).

3.6. Allowable PTE limits in octopus for human consumption

In accordance with Mexican (NOM-129-SSA1-1995 and NOM-242-SSA1-2009) and international standards (Codex Alimentarius

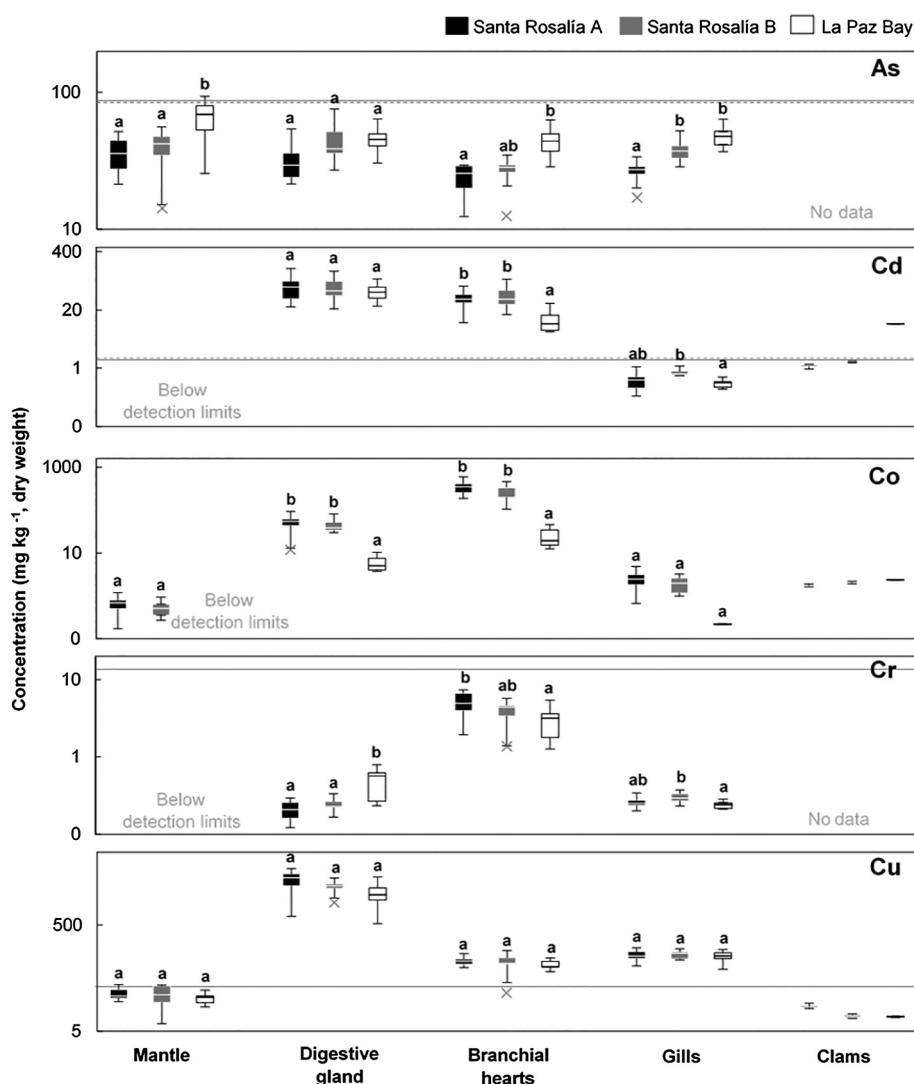


Fig. 2. Median, 25 and 75% percentiles, minimum and outliers (x) of the concentrations (note log scale) of As, Cd, Co, Cr and Cu (mg kg^{-1} , dry weight) in mantle, digestive gland, branchial hearts and gills of *Octopus hubbsorum* and whole body of the clam *Megapitaria squalida* (clams) from three locations: Santa Rosalia A and B, and La Paz Bay. Letters (a, ab, b) indicate the groups with significant differences ($p < 0.05$) between locations for each tissue. Horizontal lines indicate the standards for human consumption, according to international standards CODEX, EC and FDA (solid line) and to the Mexican government (dashed line).

Commission, European Community and US-Food and Drug Administration) (Table 6), average As, Cd, Cr, Cu, Hg, Ni and Pb concentrations in *O. hubbsorum* mantle from Santa Rosalia A and B, and La Paz Bay were below the maximum allowable limits for human consumption (Figs. 2 and 3).

4. Discussion

4.1. PTE concentrations in *O. hubbsorum*

PTE concentrations in tissues of *O. hubbsorum* from these sites in the Gulf of California were within the ranges reported for other octopus species in the world (Table 3). The only exceptions were Co, Fe and Mn in the digestive gland and branchial hearts, mainly in Santa Rosalia specimens, which exceeded by up to 100-fold (Co) the maximum values reported for *O. vulgaris* (e.g. Soldevilla, 1987; Nessim and Riad, 2003). In contrast, Co, Cr and Ni in the mantle of *O. hubbsorum* showed concentrations up to 48 times (Cr) lower than the minimum levels reported for *O. vulgaris* (Nessim and Riad, 2003; Napoleão et al., 2005).

In the current study, the elements with the highest concentrations in *O. hubbsorum* were $\text{Mg} > \text{Cu} > \text{Fe} > \text{Zn}$. This finding is consistent with essential elements that are abundant in marine organisms (Jakimska et al., 2011), including octopuses like *O. vulgaris* (e.g. Raimundo et al., 2004; Napoleão et al., 2005; Seixas et al., 2005a; Raimundo et al., 2010a). For Mg, as well as for Al, La, Sr and Tl, no

comparisons could be made because this is the first report on their concentrations in octopus.

With respect to sex, the bioaccumulation of PTEs in *O. hubbsorum* was similar in tissues of males and females, which is consistent with the findings reported for *E. cirrhosa* and *O. vulgaris* (e.g. Raimundo et al., 2004; Seixas et al., 2005b; Raimundo and Vale, 2008; Rjeibi et al., 2014).

4.2. Differences in PTE bioaccumulation between tissues

In *O. hubbsorum*, the higher concentrations of $\text{Cu} > \text{Zn} > \text{Cd} > \text{Pb} > \text{Hg}$ in the digestive gland versus other tissues were consistent with those reported for *O. vulgaris* (e.g. Raimundo et al., 2004; Seixas et al., 2005b; Karim et al., 2016), and supports the observation that this organ is the primary site for final storage and excretion of PTEs in bivalves, gastropods (Marigómez et al., 2002), and cephalopods (Miramand and Guary, 1980), since food is considered as the main source of PTEs in these organisms (Bustamante et al., 2000).

According to the classification of Miramand and Bentley (1992), Cd and Cu are the elements with the highest concentrations in the digestive gland of *O. hubbsorum*, followed by Zn, Hg and Pb. This rank in concentrations is in accordance with the concentrations reported for the digestive gland of *O. vulgaris* (Raimundo et al., 2004; Raimundo and Vale, 2008), and is related to the function of chelating agents in this organ, especially cytosolic proteins (Penicaud et al., 2017). In

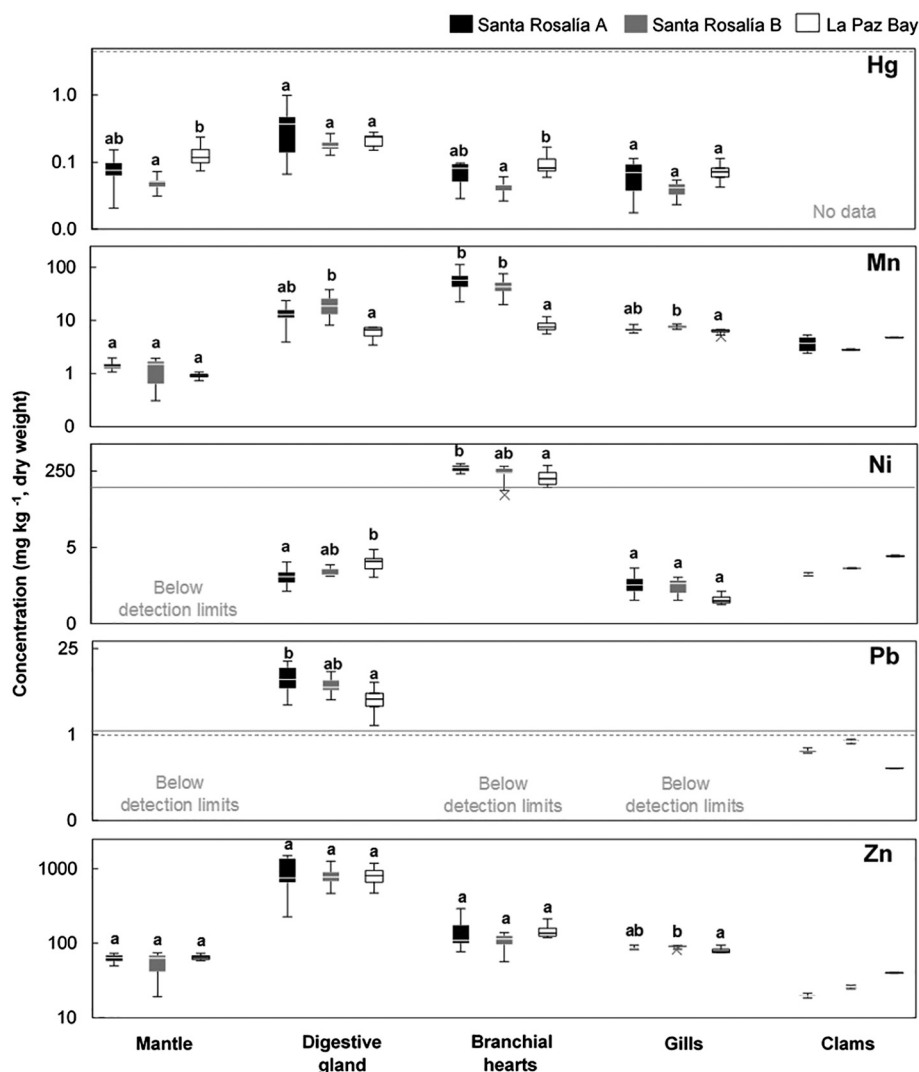


Fig. 3. Median, 25 and 75% percentiles, minimum, maximum and outliers (x) of the concentrations (note log scale) of Hg, Mn, Ni, Pb and Zn (mg kg^{-1} , dry weight) in mantle, digestive gland, branchial hearts and gills of *Octopus hubbsorum* and whole body of the clam *Megapitaria squalida* (clams) from three locations: Santa Rosalía A and B, and La Paz Bay. Letters (a, ab, b) indicate the groups with significant differences ($p < 0.05$) between locations for each tissue. Horizontal lines indicate the standards for human consumption, according to international standards CODEX, EC and FDA (solid line) and to the Mexican government (dashed line).

Table 4
Biomagnification factors of potentially toxic elements in mantle, digestive gland, and branchial hearts of *Octopus hubbsorum* in relation to the total concentration in whole clams *Megapitaria squalida*. All data are from two sites near Santa Rosalía and one in La Paz Bay, Gulf of California. The highest values are highlighted in bold.

Tissue	Locality	Cd	Co	Cu	Fe	Mn	Ni	Pb	Zn
Mantle	Santa Rosalía A	0	0.4	1.7	0.1	0.4	0	0	3.2
	Santa Rosalía B	0	0.3	2.6	0	0.4	0	0	2.1
	La Paz Bay	0	0	2.2	0	0.2	0	0	1.6
Digestive gland	Santa Rosalía A	73	31	250	2.3	4	1	30	46
	Santa Rosalía B	50	22	300	1.8	7.2	0.8	13	32
	La Paz Bay	5.4	2.5	225	2.2	1.3	0.8	24	20
Branchial hearts	Santa Rosalía A	38	203	7.2	11	15	233	0	7.7
	Santa Rosalía B	32	137	10	3.1	15	132	0	4.2
	La Paz Bay	1.3	10	9.6	3.0	2.4	57	0	4.0

cephalopods, Zn, Cu and Cd are associated with both high and low molecular weight proteins in digestive gland while Pb is associated with unknown high molecular weight proteins (Bustamante et al., 2006; Raimundo et al., 2010b). Cytosolic proteins include metallothioneins, which are known to bind Cd, Cu, Hg and Zn (Marigómez et al., 2002). Furthermore, the high Cu and Zn levels are related to their role as essential elements involved in various metabolic processes, such as the formation of metallo-dependent enzymes involved in digestion (Craig and Overnell, 2003). For its part, Pb was in lower concentrations than

Cu, Cd and Zn and this could be due to its possible elimination through other excretion pathways (nephrons, heart and intestine), as has been described for molluscs (Marigómez et al., 2002).

The high concentrations of $\text{Co} > \text{Ni} > \text{Mn} > \text{Sr} > \text{Al} > \text{V} > \text{Cr}$ in the branchial hearts of *O. hubbsorum* from these Gulf of California sites relative to all other tissues confirms that, although this organ represents a mere 0.2% of total weight, it is nonetheless an important site for the storage and clearance of elements (Mangold et al., 1989; Miramand and Fowler, 1998). The bioaccumulation of these specific elements in branchial hearts of octopuses is likely because 37% of the dry weight of this organ is composed of adenochromes (Fox, 1976), which are pigments with a high affinity for Fe (Ito et al., 1976), Co (Nakahara and Shimizu, 1985) and V (Miramand and Guary, 1980), and possibly other elements, and are considered as key molecules for their storage and clearance. Other elements, such as Al and Sr, may be up taken in gills, as has been reported in fishes (Schiffman, 1961; Exley et al., 1991), and retained by the branchial hearts through the filtration of blood from the gills (Mangold et al., 1989); however, the role of branchial hearts in storage and clearance of Al and Sr remains unknown. Along these lines, the fact that the concentrations of PTEs in *O. hubbsorum* gills were not higher relative to other tissues suggests that the former accumulates elements over the short term, as reported in bivalve molluscs such as *Ruditapes decussatus* (Bebiano and Serafim, 2003) and *Cerastoderma glaucum* (Machreki-Ajmi et al., 2008).

The low PTE concentrations in the mantle of *O. hubbsorum* relative to all other tissues is consistent with the findings reported for *O. vulgaris*

Table 5
Concentrations of potentially toxic elements (mean \pm SD) in some bivalves that represent potential prey of *Octopus hubbsorum* in Santa Rosalía and La Paz Bay, Gulf of California.

Potential preys	Locality	Cd	Co	Cu	Fe	Mn	Ni	Pb	Zn	Reference
<i>Megapitaria squalida</i>	Santa Rosalía A	1.1 \pm 0.1	1.7 \pm 0.1	14.6 \pm 1.8	122 \pm 14	3.7 \pm 1.4	1.26 \pm 0.09	0.24 \pm 0.03	19 \pm 1.2	Present study
	Santa Rosalía B	1.3 \pm 0.06	2.0 \pm 0.1	9.5 \pm 0.9	193 \pm 5	2.8 \pm 0.06	1.7 \pm 0.04	0.3 \pm 0.03	25 \pm 1.4	
	La Paz Bay	9.7 \pm 0.1	2.3 \pm 0.06	9.3 \pm 0.4	205 \pm 3	4.7 \pm 0.1	3.2 \pm 0.2	0.1 \pm 0.01	40.1 \pm 0.9	
<i>Mytilus edulis</i>	Santa Rosalía A	4.0 \pm 0.4		49.6 \pm 9.8	369 \pm 125	0.5 \pm 0.1	7.8 \pm 0.6	5.8 \pm 0.4	47.9 \pm 3.4	Cadena-Cárdenas et al., 2009
	Santa Rosalía B	3.5 \pm 0.2		17.1 \pm 3.8	150 \pm 24	0.8 \pm 0.2	6.4 \pm 0.7	4.8 \pm 0.5	52.9 \pm 4.8	
	Santa Rosalía	15 ^a 20 ^b		58.6 ^a 89 ^b		500 ^a 900 ^b			160 ^a 320 ^b	

* Average concentration by total length: a = large clams (98.6 mm), b = medium clams (75.5 mm).

Table 6
International standards for some potentially toxic elements (As, Cd, Cr, Cu, Hg, Ni and Pb) in edible tissues of fish, molluscs, bivalves and cephalopods for human consumption.

Element	Organisms	Standards (mg kg ⁻¹ , dry weight)	Reference ^a
As	Molluscs	86.0	FDA
		80.0	NOM 2009
Cd	Fish	1.3	EC
		2.0	NOM 2009
	Molluscs	4.0	FDA
		5.0	EC
	Bivalves	2.0	CODEX
		2.5	NOM 1995
Cephalopods	2.0	CODEX	
	13.0	FDA	
Cr	Molluscs	13.0	FDA
Cu	Bivalves	32.5	FDA
Hg	Fish	5.0	EC
		5.0	NOM 1995
Ni	Molluscs	80	FDA
		2.5	NOM 2009
Pb	Fish	1.5	EC
		1.5	CODEX
	Molluscs	1.7	FDA
		1.0	NOM 2009
	Bivalves	1.5	EC
		1.0	NOM 1995
	Cephalopods	1.0	EC
		1.0	EC

^a CODEX: CODEX STAN 193-1995 Amended in 2015 (CODEX FAO/WHO, 2015).

EC: European Commission (EC) No 1881/2006 (Commission, 2006).

FDA: US FDA Guidance documents (US-FDA, 2007).

NOM 1995: NOM-129-SSA1-1995 (DOF, 1996).

NOM 2009: NOM-242-SSA1-2009 (DOF, 2011)

(Miramand and Guary, 1980; Napoleão et al., 2005; Semedo et al., 2012) and other cephalopods such as *Sepia officinalis* and *Loligo vulgaris* (Rjeibi et al., 2014). This is likely because the muscle tissues are not active sites of biotransformation and accumulation of PTEs (Elnabris et al., 2013). The mantle cannot store PTEs over the long term because it lacks of specific binding proteins (e.g. metallothioneins). In addition, some elements are transferred from the mantle through the blood stream to the branchial hearts and digestive gland (Bustamante et al., 2002).

4.3. Differences in PTE bioaccumulation between sites

The concentration of PTEs in *O. hubbsorum* tissues were similar in Santa Rosalía sites A and B, despite the different contamination levels recorded in the sediments (Shumilin et al., 2011, 2013). This may be attributed to the fact that octopuses are constantly moving between areas to find food and shelter according to predator pressure (Mather, 1982). For example, *O. vulgaris* can travel up to 50 km in 40 days in search of its preferred prey (Itami, 1964), while the distance between the Santa Rosalía sites is just 2 km.

When compared to La Paz Bay, octopus tissues from Santa Rosalía were higher in almost all of the same elements found at higher concentrations in the sediments from this port, namely Mn > Cu Zn > Pb > Co > U > Cd (Shumilin et al., 2013); the exceptions were Cu and U, which occurred at similar concentrations in all octopus tissues in Santa Rosalía and in La Paz Bay, and Zn, which only showed higher concentrations gills of octopus from Santa Rosalía.

However, significant differences in elemental concentrations were found in all tissues except the mantle (Cd, Co, Cr, Mn and Ni in branchial hearts; Cd, Cr, Mn and Zn in gills; Co, Mn, and Pb in digestive gland). This highlights the high PTE clearance capacity in octopus, because even when these organisms are exposed to elevated PTEs, their concentrations in muscle tissue (mantle) are still low as has been reported in other octopus species (Napoleão et al., 2005; Raimundo and Vale, 2008).

Mn is an essential element (Newman, 2014) commonly found at high concentrations in the digestive gland and branchial hearts of octopus species (Napoleão et al., 2005). However, in *O. hubbsorum* from Santa Rosalía, Mn levels were up to five times higher than octopus from La Paz Bay in those same tissues, and almost 10 times higher than those reported for other octopus species (Table 3). Similarly, the concentrations of Co found in branchial hearts of *O. hubbsorum* from Santa Rosalía exceeded -by up to 50 times- the levels recorded for this organ in other species of octopus, squids and in bivalve tissues (Yoshida, 1981; Nessim and Riad, 2003). This is further evidence of the effect of mining wastes in the study area.

The octopuses from Santa Rosalía also showed an additional physiological compensation related to mining wastes, because in this site they showed higher concentrations of Co and Cd in branchial hearts than in their digestive glands, even when it is known that the storage and clearance of these elements take place primarily in the digestive gland (Miramand and Bentley, 1992). This suggests that the uptake capacity of the digestive gland was exceeded, and that hearts serve as ancillary clearance organs (Bustamante et al., 1998b).

Although Cu, Zn and U are also among the most concentrated elements in Santa Rosalía sediments, their concentrations in tissues of *O. hubbsorum* did not reflect this pollution. This can be attributable to the fact that Cu and Zn are essential elements whose concentrations are regulated by various homeostatic mechanisms. Both elements are involved in the synthesis of several metallo-dependent enzymes (Craig and Overnell, 2003). Cu is mainly used in the synthesis of hemocyanin, a respiratory pigment containing 0.25% copper (Bustamante et al., 2000), as well as in the production of amine oxidase, a toxic component present in the salivary glands of octopuses (Nessim and Riad, 2003; Raimundo and Vale, 2008). In addition, it has been reported that many species of crustaceans and molluscs maintain a relatively constant total body load of Cu despite wide environmental concentrations (Taylor and Anstiss, 1999) and this could be the case for *O. hubbsorum*. However, to be able to better explain this exception, studies on the physiology of Cu in this species would be necessary. For its part, the lack of differences in

U concentrations between octopuses from Santa Rosalia and La Paz Bay suggests that this element is available at similar concentrations in both sites, possibly due to the natural U deposits in La Paz (SGM, 2014).

Although La Paz Bay is not considered an area highly polluted by PTEs (Rodríguez-Castañeda et al., 2006), octopuses from this site had higher Ag, Cr, Ni and Sr in the digestive glands and higher As and Hg in all tissues when compared to results from Santa Rosalia. This finding confirms that As clearance does not occur mainly in the digestive gland of cephalopods and it has a high affinity for muscle tissue in the mantle and arms (e.g. Napoleão et al., 2005; Seixas et al., 2005b; Semedo et al., 2012). However, the high As in octopus tissues from La Paz Bay were in contrast to the low As in sediments and algae from this location (Rodríguez-Castañeda et al., 2006). A plausible explanation involves the diet, because if octopuses from La Paz Bay consume larger quantities of crustaceans relative to those from Santa Rosalia, this could lead to higher As concentrations in the tissues of the former, as has been suggested for *N. macromphalus* (Bustamante et al., 2000).

With respect to Hg, octopus accumulates this element primarily from the diet, so it is stored and cleared by the digestive gland (Seixas et al., 2005b; Rjeibi et al., 2014). Foodborne Hg in cuttlefish, squid (Rjeibi et al., 2014) and bivalves (Marigómez et al., 2002) is commonly transferred through the digestive gland and stored in muscle tissue; however, in some bivalves aqueous Hg accumulates both in the mantle epithelial cells and in gills (Marigómez et al., 2002). Accordingly, the high Hg concentrations in the mantle and branchial hearts of *O. hubbsorum* from La Paz Bay could be due to high Hg concentrations in water and its direct uptake through the epithelium and gills, respectively. However, there are no studies available reporting aqueous Hg concentrations in La Paz Bay.

4.4. Biomagnification of PTEs

At all sites, *O. hubbsorum* had higher concentrations of Cd, Co, Cu, Fe, Mn, Ni, Pb and Zn in its tissues relative to the concentrations of the same elements in one of its prey (the clam *M. squallida*), indicating biomagnification. As in other octopus species, *O. hubbsorum* feed upon bivalve molluscs (López-Uriarte et al., 2010), which usually have high PTE concentrations due to their direct exposure to the substrate, filter-feeding, and tolerance to high pollutant levels, among others (Jakimska et al., 2011). However, molluscs account for just 30% of the octopus diet (López-Uriarte et al., 2010) therefore future research should consider other prey such as crustaceans (\pm 60% of the diet) (López-Uriarte et al., 2010) for a more comprehensive assessment of PTE biomagnification in this species.

4.5. PTEs standards for human consumption of octopus

The mean concentrations of PTEs in the mantle of *O. hubbsorum* collected in Santa Rosalia and La Paz Bay do not pose a public-health risk, according to international standards for human consumption (Table 6). Among the tissues examined, only the mantle is consumed frequently by humans. However, the digestive gland and/or branchial hearts of octopus from Santa Rosalia or La Paz Bay exceeded the Cd, Cu, Ni and Pb levels established in the standards mentioned above (Figs. 2 and 3); hence these organs are not suitable for consumption.

5. Conclusions

This study presents the first data on PTEs in octopuses from the Gulf of California, and provides evidence of high bioaccumulation and biomagnification of PTEs in octopuses from sites known to have historical metal contamination. Nevertheless, *O. hubbsorum* show the ability to regulate the high concentrations of most PTEs through specific organs (digestive gland and branchial hearts) maintaining similar elemental composition of the mantle among sites and levels below PTE standards for human consumption of octopus.

Acknowledgments

This research was made possible through funding from the Natural Sciences and Engineering Research Council of Canada Discovery (950-230607) program, the Canada Research Chairs (4293899-2012 and 312237-2012) program and the projects 1535, 1698 and 20171428 of the Secretaría de Investigación y Posgrado (SIP). Nefertiti Taydé Roldán Wong is a fellow student of BEIFI and CONACyT, the results presented here are part of her M. Sc. thesis. Ana Judith Marmolejo Rodríguez, Evgueni Shumilin, Bertha Patricia Ceballos Vázquez and Marcial Arellano Martínez received grants from COFAA, EDI, and SNI-CONACyT. We thank María Elena Sánchez-Salazar, M. Sc. for her editorial contribution to the English manuscript, and Angella Mercer, PTEch, for her help with the laboratory analysis.

References

- Armendáriz-Villegas, E.J., Ceballos-Vázquez, B.P., Markaida, U., Arbitia-Cárdenas, A., Medina-López, M.A., Arellano-Martínez, M., 2014. Diet of *Octopus bimaculatus* Verrill, 1883 (Cephalopoda: Octopodidae) in Bahía De Los Ángeles, Gulf of California. *J. Shellfish Res.* 33 (1), 305–314. <http://dx.doi.org/10.2983/035.033.0129>.
- Bebiano, M.J., Serafim, M.A., 2003. Variation of metal and metallothionein concentrations in a natural population of *Ruditapes decussatus*. *Arch. Environ. Contam. Toxicol.* 44, 53–66. <http://dx.doi.org/10.1007/s00244-002-2004-7>.
- Bustamante, P., Caurant, F., Fowler, S.W., Miramand, P., 1998a. Cephalopods as a vector for the transfer of cadmium to top marine predators in the north-east Atlantic Ocean. *Sci. Total Environ.* 220, 71–80. [http://dx.doi.org/10.1016/S0048-9697\(98\)00250-2](http://dx.doi.org/10.1016/S0048-9697(98)00250-2).
- Bustamante, P., Cherel, Y., Caurant, F., Miramand, P., 1998b. Cadmium, copper and zinc in octopuses from Kerguelen Islands, Southern Indian Ocean. *Polar Biol.* 19, 264–271. <http://dx.doi.org/10.1007/s003000050244>.
- Bustamante, P., Grigioni, S., Boucher-Rodoni, R., Caurant, F., Miramand, P., 2000. Bioaccumulation of 12 trace elements in the tissues of the nautilus *Nautilus macromphalus* from New Caledonia. *Mar. Pollut. Bull.* 40 (8), 688–696. [http://dx.doi.org/10.1016/S0025-326X\(00\)00005-9](http://dx.doi.org/10.1016/S0025-326X(00)00005-9).
- Bustamante, P., Teysié, J.-L., Fowler, S.W., Cotret, O., Danis, B., Miramand, P., Warnau, M., 2002. Biokinetics of zinc and cadmium accumulation and depuration at different stages in the life cycle of the cuttlefish *Sepia officinalis*. *Mar. Ecol. Prog. Ser.* 231, 167–177. <http://dx.doi.org/10.3354/meps231167>.
- Bustamante, P., Bertrand, M., Boucaud-Camou, E., Miramand, P., 2006. Subcellular distribution of Ag, Cd, Co., Cu, Fe, Mn, Pb and Zn in the digestive gland of the common cuttlefish *Sepia officinalis*. *J. Shellfish Res.* 25, 987–993. [http://dx.doi.org/10.2983/0730-8000\(2006\)25\[987:SDOACC\]2.0.CO;2](http://dx.doi.org/10.2983/0730-8000(2006)25[987:SDOACC]2.0.CO;2).
- Cadena-Cárdenas, L., Méndez-Rodríguez, L., Zenteno-Savín, T., García-Hernández, J., Acosta-Vargas, B., 2009. Heavy metal levels in marine molluscs from areas with, or without, mining activities along the Gulf of California, Mexico. *Arch. Environ. Contam. Toxicol.* 57, 96–102. <http://dx.doi.org/10.1007/s00244-008-9236-0>.
- CODEX, 2015. General Standard for Contaminants and Toxins in Food and Feed (CODEX STAN 193-1995). Amended in 2015. Codex Alimentarius-FAO-WHO (59 pp. online). www.fao.org.
- Commission, 2006. Commission Regulation (EC) No 1881/2006 of December 2006: Setting Maximum Levels for Certain Contaminants in Foodstuffs (26 pp. online). www.eur-lex.europa.eu.
- Craig, S., Overnell, J., 2003. Metals in squid, *Loligo forbesi*, eggs and hatchlings. No evidence for a role for Cu- or Zn-metallothionein. *Comp. Biochem. Physiol. C* 134, 311–317. [http://dx.doi.org/10.1016/S1532-0456\(02\)00274-0](http://dx.doi.org/10.1016/S1532-0456(02)00274-0).
- DOF, Diario Oficial de la Federación, 1996. Norma Oficial Mexicana NOM 1995: NOM-129-SSA1-1995. Bienes y servicios. Productos de pesca. Moluscos cefalópodos y gasterópodos. INAPESCA-SAGARPA, México (online). www.dof.gob.mx.
- DOF, Diario Oficial de la Federación, 2011. Norma Oficial Mexicana NOM 2009: NOM-242-SSA1-2009. Productos y servicios. Productos de pesca. Especificaciones sanitarias y métodos de prueba. Metales pesados. INAPESCA-SAGARPA, México (online). www.dof.gob.mx.
- Elnabris, K.J., Muzyed, S.K., El-Ashgar, N.M., 2013. Heavy metal concentrations in some commercially important fishes and their contribution to heavy metals exposure in Palestinian people of Gaza Strip (Palestine). *J. Assoc. Arab. Univ. Basic Appl. Sci.* 13, 44–51. <http://dx.doi.org/10.1016/j.jaubas.2012.06.001>.
- Exley, C., Chappell, J.S., Birchall, J.D., 1991. A mechanism for acute aluminium toxicity in fish. *J. Theor. Biol.* 151 (3), 417–428. [http://dx.doi.org/10.1016/S0022-5193\(05\)80389-3](http://dx.doi.org/10.1016/S0022-5193(05)80389-3).
- FAO, 2016. Cephalopods of the world. An annotated and illustrated catalogue of cephalopod species known to date. Octopods and Vampire Squids. In: Jereb, P., Roper, C.F.E., Norman, M.D., Finn, J.K. (Eds.), FAO Species Catalogue for Fishery Purposes. Rome, (370 pp. online). www.fao.org.
- Fox, D.L., 1976. *Animal Biochromes and Structural Colors*. University of California Press, Berkeley-Los Angeles-London (433 pp. ISBN: (9780520023475).
- Grayson, J., Sekadende, B., 2014. Assessment of heavy metal pollution in *Octopus cyanea* in the coastal waters of Tanzania. *J. Health Pollut.* 4 (6), 10–17. <http://dx.doi.org/10.5696/2156-9614-4-6.10>.
- Gutiérrez-Galindo, E., Villaescusa-Celaya, J.A., Arreola-Chimal, A., 1999. Bioaccumulation of metals in mussels from four sites of the coastal region of Baja

- California. *Cienc. Mar.* 25 (4), 557–578. <http://dx.doi.org/10.7773/cm.v25i4.726>.
- Itami, K., 1964. Marking and release study in the octopus. *Aquaculture* 12, 119–125.
- Ito, S., Nardi, G., Protta, G., 1976. Structures of adenochromines A and B, the iron (III) binding amino-acids of a unique group of peptides, adenochromes from *Octopus vulgaris*. *J. Chem. Soc. Chem. Commun.* 1042–1043. <http://dx.doi.org/10.1039/C39760001042>.
- Jakimska, A., Konieczka, P., Skóra, K., Namiesnik, J., 2011. Bioaccumulation of metals in tissues of marine animals, part II: metal concentrations in animal tissues. *Pol. J. Environ. Stud.* 20 (5), 1127–1146.
- Karim, S., Aouniti, A., Belbachir, C., Rahhou, I., El abed, S., Hammouti, B., 2016. Metallic contamination (Cd, Pb, Cu, Zn, Fe, Co) of the octopus (*Octopus vulgaris* Cuvier, on 1797) fished in the Mediterranean coast from the north east of Morocco. *J. Chem. Pharm. Res.* 8 (2), 821–828.
- López-Urriarte, E., Ríos-Jara, E., Pérez-Peña, M., 2005. Range extension for *Octopus hubbsorum* Berry 1953 (Mollusca: Octopodidae) in the Mexican Pacific. *Bull. Mar. Sci.* 7 (2), 171–175.
- López-Urriarte, E., Ríos-Jara, E., González-Rodríguez, M.E., 2010. Diet and feeding habits of *Octopus hubbsorum* Berry, 1953, in the Central Mexican Pacific. *Veliger* 51 (1), 26–42.
- Machreki-Ajmi, M., Ketata, I., Laghar-Chaabouni, R., Hamza-Chaffai, A., 2008. The effect of in situ cadmium contamination on some biomarkers in *Cerastoderma glaucum*. *Ecotoxicology* 17, 1–11. <http://dx.doi.org/10.1007/s10646-007-0166-9>.
- Mangold, K., Bidder, A.M., Boletzky, S., 1989. Appareils excréteurs et excrétion. P.p 439–457. In: Mangold, K., Grassé, P.P. (Eds.), *Traité de zoologie: anatomie, systématique, biologie*. France, (804 pp. ISBN: 2225804192 9782225804199).
- Marigómez, I., Soto, M., Cajaraville, M.P., Angulo, E., Giamberini, L., 2002. Cellular and subcellular distribution of metals in molluscs. *Microsc. Res. Tech.* 56, 358–392. <http://dx.doi.org/10.1002/jemt.10040>.
- Martin, J., Flegal, A., 1975. High copper concentrations in squid livers in association with elevated levels of silver, cadmium and zinc. *Mar. Biol.* 30, 51–55. <http://dx.doi.org/10.1007/BF00393752>.
- Mather, J.A., 1982. Factors affecting the spatial distribution of natural populations of *Octopus joubini* Robson. *Anim. Behav.* 30, 1166–1170. [http://dx.doi.org/10.1016/S0003-3472\(82\)80207-8](http://dx.doi.org/10.1016/S0003-3472(82)80207-8).
- Miramand, P., Bentley, D., 1992. Concentration and distribution of heavy metals in tissues of two cephalopods, *Eledone cirrhosa* and *Sepia officinalis*, from the French coast of the English Channel. *Mar. Biol.* 114, 407–414. <http://dx.doi.org/10.1007/BF00350031>.
- Miramand, P., Fowler, S., 1998. Bioaccumulation and transfer of vanadium in marine organisms. P.p. 167–197. In: Nriagu, J.O. (Ed.), *Vanadium in the Environment. Part 1: Chemistry and Biochemistry*. John Wiley & Sons, New York (410 pp).
- Miramand, P., Guary, J.C., 1980. High concentrations of some heavy metals in tissues of the Mediterranean octopus. *Bull. Environ. Contam. Toxicol.* 24, 783–788. <http://dx.doi.org/10.1007/BF01608189>.
- Mshana, J.G., Sekadende, B., 2014. Assessment of Heavy Metal Pollution in *Octopus cyanea* in the Coastal Waters of Tanzania. *J. Health Pollut.* 4 (6), 10–17. <http://dx.doi.org/10.5696/2156-9614-4-6-10>.
- Muñoz-Barbosa, A., Huerta-Díaz, M.A., 2013. Trace metal enrichments in nearshore sediments and accumulation in mussels (*Modiolus capax*) along the eastern coast of Baja California, Mexico: environmental status in 1995. *Mar. Pollut. Bull.* 77, 71–81. <http://dx.doi.org/10.1016/j.marpolbul.2013.10.030>.
- Nakahara, M., Shimizu, C., 1985. Cobalt-binding substances in the branchial heart of *Octopus vulgaris*. *Bull. Jpn. Soc. Sci. Fish.* 51 (7), 1195–1199. <http://dx.doi.org/10.2331/suisan.51.1195>.
- Napoleão, P., Pinheiro, T., Sousa Reis, C., 2005. Elemental characterization of tissues of *Octopus vulgaris* along the Portuguese coast. *Sci. Total Environ.* 345, 41–49. <http://dx.doi.org/10.1016/j.scitotenv.2004.10.026>.
- Nessim, R.B., Riad, R., 2003. Bioaccumulation of heavy metals in *Octopus vulgaris* from coastal waters of Alexandria (Eastern Mediterranean). *Chem. Ecol.* 19 (4), 275–281. <http://dx.doi.org/10.1080/02757540310001595907>.
- Newman, M.C., 2014. *Fundamentals of Ecotoxicology: The Science of Pollution*, Fourth Edition. CRC Press, New York (633 pp. ISBN-10: 1466582294).
- Penicaud, V., Lacoue-Labarthe, T., Bustamante, P., 2017. Metal bioaccumulation and detoxification processes in cephalopods: a review. *Environ. Res.* 155, 123–133. <http://dx.doi.org/10.1016/j.envres.2017.02.003>.
- Raimundo, J., Vale, C., 2008. Partitioning of Fe, Cu, Zn, Cd, and Pb concentrations among eleven tissues of *Octopus vulgaris* from the Portuguese coast. *Cienc. Mar.* 34 (3), 297–305. <http://dx.doi.org/10.7773/cm.v34i3.1402>.
- Raimundo, J., Caetano, M., Vale, C., 2004. Geographical variation and partition of metals in tissues of *Octopus vulgaris* along the Portuguese coast. *Sci. Total Environ.* 325, 71–81.
- Raimundo, J., Pereira, P., Vale, C., Caetano, M., 2005. Fe, Zn, Cu and Cd concentrations in the digestive gland and muscle tissues of *Octopus vulgaris* and *Sepia officinalis* from two coastal areas in Portugal. *Cienc. Mar.* 31 (1B), 243–251. <http://dx.doi.org/10.7773/cm.v31i12.91>.
- Raimundo, J., Costa, P.M., Vale, C., Costa, M.H., Moura, I., 2010a. Metallothioneins and trace elements in digestive gland, gills, kidney and gonads of *Octopus vulgaris*. *Comp. Biochem. Physiol. C* 152, 139–146. <http://dx.doi.org/10.1016/j.cbpc.2010.03.009>.
- Raimundo, J., Vale, C., Duarte, R., Moura, I., 2010b. Association of Zn, Cu, Cd and Pb with protein fractions and sub-cellular partitioning in the digestive gland of *Octopus vulgaris* living in habitats with different metal levels. *Chemosphere* 81, 1314–1319. <http://dx.doi.org/10.1016/j.chemosphere.2010.08.029>.
- Rjeibi, M., Metian, M., Hajji, T., Guyot, T., Chaouacha-Chekir, R.B., Bustamante, P., 2014. Interspecific and geographical variations of trace metal concentrations in cephalopods from Tunisian waters. *Environ. Monit. Assess.* 186 (6), 3767–3783. <http://dx.doi.org/10.1007/s10661-014-3656-2>.
- Rodríguez-Castañeda, A.P., Sánchez-Rodríguez, I., Shumilin, E.N., Sapozhnikov, D., 2006. Element concentrations in some species of seaweeds from La Paz Bay and La Paz Lagoon, south-western Baja California, Mexico. *J. Appl. Phycol.* 18, 399–408. <http://dx.doi.org/10.1007/s10811-006-9040-z>.
- Rossi, A., Pellegrini, D., Belcarì, P., Barghigiani, C., 1993. Mercury in *Eledone cirrhosa* from the Northern Tyrrhenian Sea: contents and relations with life cycle. *Mar. Pollut. Bull.* 26 (12), 683–686. [http://dx.doi.org/10.1016/0025-326X\(93\)90551-T](http://dx.doi.org/10.1016/0025-326X(93)90551-T).
- Schiffman, R., 1961. A perfusion study of the movement of strontium across the gills of rainbow trout (*Salmo gairdnerii*). *Biol. Bull.* 120 (1), 110–117. <http://dx.doi.org/10.2307/1539341>.
- Seixas, S., Pierce, G.J., 2005. Vanadium, rubidium and potassium in *Octopus vulgaris* (Mollusca: Cephalopoda). *Sci. Mar.* 69 (2), 215–222. <http://dx.doi.org/10.3989/scimar.2005.69n2215>.
- Seixas, S., Bustamante, P., Pierce, G.J., 2005a. Interannual patterns of variation in concentrations of trace elements in arms of *Octopus vulgaris*. *Chemosphere* 59, 1113–1124. <http://dx.doi.org/10.1016/j.chemosphere.2004.11.099>.
- Seixas, S., Bustamante, P., Pierce, G.J., 2005b. Accumulation of mercury in the tissues of the common octopus *Octopus vulgaris* (L.) in two localities on the Portuguese coast. *Sci. Total Environ.* 340, 113–122. <http://dx.doi.org/10.1016/j.scitotenv.2004.08.012>.
- Semedo, M., Reis-Henriques, M.A., Rey-Salgueiro, L., Oliveira, M., Delerue-Matos, C., Morais, S., Ferreira, M., 2012. Metal accumulation and oxidative stress biomarkers in octopus (*Octopus vulgaris*) from Northwest Atlantic. *Sci. Total Environ.* 433, 230–237. <http://dx.doi.org/10.1016/j.scitotenv.2012.06.058>.
- SGM, Servicio Geológico Mexicano, 2014. El Uranio en México. SGM. Secretaría de Economía (online). www.sgm.gob.mx.
- Shumilin, E., Gordeev, V., Rodríguez-Figueroa, G., Demina, L., Choumilina, K., 2011. Assessment of geochemical mobility of metals in surface sediments of the Santa Rosalia mining region, western Gulf of California. *Arch. Environ. Contam. Toxicol.* 60, 8–25. <http://dx.doi.org/10.1007/s00244-010-9532-3>.
- Shumilin, E., Jiménez-Illescas, A.R., López-López, S., 2013. Anthropogenic contamination of metals in sediments of the Santa Rosalia Harbor, Baja California Peninsula. *Bull. Environ. Contam. Toxicol.* 90 (3), 333–337. <http://dx.doi.org/10.1007/s00128-012-0923-1>.
- Soldevilla, F., 1987. Metales pesados en el pulpo común (*Octopus vulgaris*) del banco sahariano (costas NO de África). *Alimentaria* 33, 33–37.
- Taylor, H.H., Anstiss, J.M., 1999. Cooper and haemocyanin dynamics in aquatic invertebrates. *Mar. Freshw. Res.* 50 (8), 907–931. <http://dx.doi.org/10.1071/MF99117>.
- US-FDA, United States Food and Drug Administration, 2007. National Shellfish Sanitation Program-Guide for the Control of Molluscan Shellfish. Section IV. Guidance Documents. Chapter II. Growing Areas (4 pp. online). www.fda.gov.
- Volke-Sepúlveda, T., Solórzano, G., Rosas, A., Izumikawa, C., Aguilar, G., Velasco, J., Flores, S., 2003. Remediación de sitios contaminados por metales provenientes de jales mineros en los distritos de El Triunfo-San Antonio y Santa Rosalia, Baja California, Sur. CENICA, México (37 pp).
- Yoshida, K., 1981. High Accumulation of Cobalt-60 by the Particular Organ or Part of Some Species in Mollusca (NIRS-M-39). 14(24). National Institute of Radiological Sciences, Chiba, Japan, pp. 207–226.